

E-ISSN: 2347-5129 P-ISSN: 2394-0506 (ICV-Poland) Impact Value: 5.62 (G1F) Impact Factor: 0.549 IJFAS 2018; 6(3): 251-259 © 2018 IJFAS www.fisheriesjournal.com Received: 01-03-2018 Accepted: 02-04-2018

Robert Ondiba

Kenya Marine and Fisheries Research Institute (KMFRI), Sangoro Station, PO Box 136-40111, Pap Onditi, Kenya

Reuben Omondi

Department of Aquatic and Fishery Sciences, Kisii University, P.O. Box 408-40200, Kisii, Kenya

Kobingi Nyakeya

Kenya Marine and Fisheries Research Institute (KMFRI), Baringo Station, PO Box 31-30406, Kampi Ya Samaki, Kenya

Jacob Abwao

Kenya Marine and Fisheries Research Institute (KMFRI), Sangoro Station, PO Box 136-40111, Pap Onditi, Kenya

Safina Musa

Kenya Marine and Fisheries Research Institute (KMFRI), Sangoro Station, PO Box 136-40111, Pap Onditi, Kenya

Elijah Oyoo-Okoth

School of Environmental Studies, Department of Environmental Biology, University of Eldoret, P.O. Box 1125-30100, Eldoret Kenya

Correspondence Robert Ondiba

Kenya Marine and Fisheries Research Institute (KMFRI), Sangoro Station, PO Box 136-40111, Pap Onditi, Kenya

Environmental constraints on macrophyte distribution and diversity in a tropical endorheic freshwater lake (Lake Baringo, Kenya)

Robert Ondiba, Reuben Omondi, Kobingi Nyakeya, Jacob Abwao, Safina Musa and Elijah Oyoo-Okoth

Abstract

The structure of macrophyte assemblages can be affected by myriad factors, including physical and chemical characteristics of the water body. However, knowledge on the environmental factors affecting macrophyte diversity in endorheic freshwater lakes is limited. In this study the patterns of plant species diversity and composition and their potential determinants in Lake Baringo, Kenya, is described. Macrophyte sampling in Lake Baringo was done monthly from January 2015 to April 2016 using quadrats (1 m \times 1 m) placed along transects perpendicular to the shoreline. Water temperature, pH, conductivity, dissolved oxygen (DO), salinity and alkalinity were measured *in situ* at each of the sampling sites. Findings revealed that macrophyte species composition and assemblage exhibited significant spatial differences (P<0.05), where areas near river inlets had higher species composition and percentage cover. The findings improve the understanding of floristic patterns and plant biodiversity in the lake.

Keywords: Macrophyte assemblages, physical chemical variables, floristic patterns, rift valley, limnology

1. Introduction

Lakes offer a wide variety of ecological services, including water supply, nutrient retention [1] and habitats for many aquatic plants and animals, which interact as a balanced ecosystem. Endorheic lakes, in particular, are known to be vulnerable ecosystems through time, because their hydrological budget is mostly ruled by evaporation due to the absence of a surficial drainage output. As a result, endorheic lakes are very sensitive to changes in air temperature and precipitation and thus they deserve special attention in the on-going debate about the possible effects of environmental change on biodiversity. The aquatic macrophytes including macroalgae of the divisions Chlorophyta (green algae), Xanthophyta (yellow-green algae), Rhodophyta (red algae), the "blue-green algae" (Cyanobacteria), Bryophyta (mosses and liverworts), Pteridophyta (ferns) and Spermatophyta (seed-bearing plants) are in intimate contact with the lake environment because their roots are either in the sediment or immersed/floating in the water [2-4]. These plants include emergent macrophytes (plants that are rooted in submersed soils or soils that are periodically inundated), floating-leaved macrophytes (plants rooted to the lake bottom with leaves that float on the surface of the water), submersed macrophytes (plants that grow completely submerged under the water, with roots or rootanalogues in, attached to, or closely associated with the substrate) and free-floating macrophytes (plants that typically float on or under the water surface) [2, 5, 6]. Ecological importance of aquatic macrophytes include: provision of energy to herbivore and detritivore food webs [7], influencing the physical and chemical conditions of the water column [8, 9] and nutrient cycling [10, 11]. Aquatic macrophytes serve as a base of aquatic food-chains and therefore they aggressively contribute to the promotion and maintenance of food webs and services in freshwater ecosystems [12]. Their population response to environmental changes renders them important bioindicators of environmental conditions and long-term ecological changes in water quality [13-16].

Distinctive features of macrophyte populations in aquatic ecosystems are defined by their structural attributes such as species composition, the types and distribution of different growth

forms, abundance and diversity. These attributes respond to factors such as the physical and chemical characteristics of the water body [17-23], environmental factors [21, 24-26] biological factors [27-30], hydrological regime [31-34] as well as suits of human activities [35-38].

Factors potentially influencing macrophyte community distribution and variation in freshwater systems have been considered at various scales [5, 21, 24, 32, 39]. First, there is the large, regional scale [32, 40-42] where these community characteristics are usually primarily driven by geographyrelated factors (e.g. temperate versus tropical climate). Second, is medium or catchment scale, where, for example, hydrological and chemical variation in the system may be important [21, 43, 44]. Third, is small scale, related to environmental features of specific habitats and communities, and the biological interactions at this level [24, 45]. A number of studies have investigated the effects of environmental variables on the macrophyte distribution in many forms of freshwater lakes, but little is known about the dynamics of these communities in freshwater endorheic lakes. In this study, the general patterns of aquatic macrophyte diversity in Lake Baringo, were described. Firstly, the species richness and life forms of aquatic macrophytes at different parts of the lake was examined. Further, the changes in vegetation composition relative to environmental changes were analyzed. Lake Baringo, a freshwater lake in the Kenyan Rift Valley, is fed by perennial and ephemeral rivers, direct rainfall, and hot springs within Ol Kokwe Island, near the centre of the lake. The lake has no surface outlet and despite high evaporation rates it maintains fresh waters [46]. The lake faces humaninduced changes as a result of land- and water-use [47-48]. The perturbations include: poor agricultural systems on the catchment that lead to soil erosion, changes in hydrology due to water abstraction for horticulture, domestic use, and industrial use. Although significant efforts have been made, many restoration programs have failed because of a lack of knowledge about the crucial environmental factors and how they regulate the macrophyte community in the lake. In view of the significant role played by macrophytes in freshwater ecosystems, understanding and quantifying the environmental factors that influence their distribution patterns, will improve management practices in this wetland of international importance.

2. Materials and Methods

2.1 Study area

Lake Baringo is located between latitude 0°30'N and 0°45'N and longitude 36°00'E and 36°10'E (Fig. 1) and lies approximately 60 km north of the equator at an altitude of 900 metres above sea level. It has a mean depth of 4.7 m with the deepest point being about 11.0 m at high water levels [49]. Other morphometric and hydrological characteristics of Lake Baringo are summarized in Table 1. The lake area has two rainy seasons and a mean annual rainfall of 635 mm. The surface area covers slightly over 130 km² [50]. The catchment area which is about 6820 km² includes a large part of the western escarpment of the Rift Valley, where most of the water is derived from. In the lake are five major islands, the biggest being the volcanic Kokwa Island, from which a number of hot-springs discharge into the lake. In this study, the lake was apportioned into three ecological zones based on earlier studies, the southern, central and northern. The southern zone drains rivers Molo, Endau, Perkerra, and Ol-Arabel. The central zone has only River Mukutan draining its waters into the lake via the eastern side. The northern zone is characterized by stony substrata, but the deepest part of the

2.2 Sampling sites

A total of nine sampling sites were established for the current surveys. Thus, in the southern zone there were three sites: Salabani (S1), Ngambo (S2) and Kiserian (S3). In the Central zone the sites were BMU (C1), Kokwa Island (C2) and Long'icharo (C3); whereas the three sites at the northern side include Katuuit (N1), Loruk (N2) and Komolion (N3).

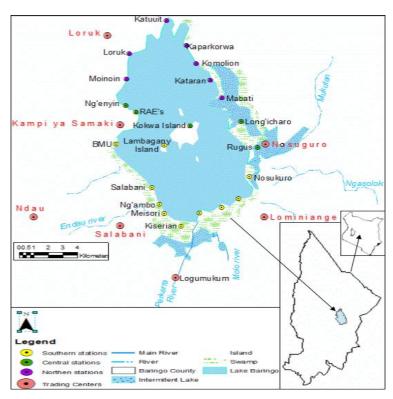


Fig 1: Map of Lake Baringo showing macrophyte study stations

Table 1: Mean morphometric characteristics of Lake Baringo, during the study period

Feature	Unit	Measures				
Surface area	km ²	130				
Lakeshore length	km	24				
Lake width	km	12				
Catchment area	km ²	6820				
Precipitation	Mm year-1	700				
Water volume	m^3	825×10^{6}				
Mean ambient temperature	°C	26.2				
Rainfall	mm	500-1100				
Basin discharge	$m^3 s^{-1}$	180-250				
Evaporation	$m^3 s^{-1}$	1500-2000				
Underground seepage	$m^3 s^{-1}$	50-150				

2.3 Sampling for aquatic macrophytes

Macrophyte sampling in Lake Baringo was done monthly from January 2015 to April 2016. The macrophytes were sampled using quadrats (1 m × 1 m) that were placed along transects perpendicular to the shoreline. The number of transects in each sampling site ranged from one to five. Within each transect, the distance between successive quadrats was constant, and the number of quadrats per transect varied from five to sixteen. This procedure takes into account habitat area and the size of the macrophyte stands, ensuring a representative inventory of aquatic flora. In addition, this procedure accounts for the variation of spatial distribution in the margins, allowing the investigation of associations between macrophyte assemblage attributes and environmental variables. Inside each 1 m × 1 m quadrat, macrophyte species were visually scored for percent cover to provide an estimate of abundance. Rakes were used to sample the submersed macrophytes. The collected specimen were transported to the University of Eldoret, where the plants were identified according to specified key and illustrations [51]. Despite some taxa not identified to species level, for the sake of simplicity, the term species was used to represent taxonomic units.

2.4 Determination of environmental parameters

Environmental parameters, namely water temperature, pH, conductivity, dissolved oxygen (DO) concentration, salinity and alkalinity were measured *in situ* at each of the sampling sites, using a calibrated JENWAY 3405 electrochemical analyzer (Barloword Scientific Ltd, Essex, UK), with independent probes for each variable. Turbidity was measured using turbidity meter (Thermo Scientific™ AQ4500 Turbidity Meter, UK). Nitrates, dissolved organic phosphates and dissolved organic carbon followed protocols in APHA 2005 ^[52].

2.5. Data analysis

Physico-chemical parameters were presented as means \pm SEM. Spatial differences in the values of the physico-chemical parameters were determined using One Way ANOVA. Vegetation distribution was determined as median cover among the sites sampled during the study period. Spatial differences, in the cover of the plant species was analyzed using Kruskall-Wallis test. A sites \times vegetation species matrix was constructed during comparison of species attributes among sites [53] before carrying out a hierarchical cluster analysis. An average linkage method was selected with the cophenetic correlation criterion and the optimum number of clusters (k). Based on the macrophyte measures, the similarities of the measured parameters were compared

among sites using exploratory cluster analysis. The dichotomous classification technique expressed the measured parameters in an ordered table, constructed from site-variable matrix. The outputs are viewed as dendrograms that illustrate sampling sites exhibiting similar species composition. For ease of comparison, the scale was reduced to percentage by dlink/dmax*100. Principal Component Analysis (PCA) was used for determining the relationships between the individual vegetation species attributes and the measured physical and chemical (environmental) variables [54]. PCA assigns a loading value to each variable on each factor (principal component, PC) and the same assignment is given to the scores (environmental variables). The multicollinearity for PCA was determined using variance decomposition proportions [55]. The software STATISTICA 10.0 (ver. 10) (Stat Soft, Inc., Tulsa, OK) was employed in all the analysis including the multivariate statistical analysis.

3. Results

An overview of the physical and chemical variables in Lake Baringo observed at the inlet sites (S1, S2, S3), central sites (C1, C2, and C3) and northern sites (N1, N2 and N3) are shown in Table 2. All the physical and chemical parameters demonstrated significant (P < 0.05) spatial variations. Sites S1, S2 and S3 located near inlets exhibited significantly lower temperature, pH, conductivity, salinity, alkalinity and DOC but had higher turbidity, TDS, DO, NO₃-N and DOP. The northern sites had significantly (P < 0.05) higher temperature, pH, conductivity and alkalinity.

A total of 30 plant species belonging to 16 families were identified at the nine sampling sites in Lake Baringo between January and December 2015 (Table 3). Overall, the family Poaceae had the highest number of species (five) followed by Cyperaceae with four species whereas Pappilionaceae and Onagraceae had three species each. A total of seven families had a single species each. In terms of life forms, 67% of the plant species were emergents followed by free floating type (17%). The submerged macrophyte type was represented by four species belonging to three families, whereas the floating rooted type had the least representation by species. It was observed that the southern zone of Lake Baringo was relatively rich in macrophyte species (29 species) compared with the central (22 species) and northern zones (21 species). The temporal differences in the mean [range] of macrophyte species in Lake Baringo is also presented in Table 3. There were significant spatial differences in the percent vegetation cover among sites (P<0.05) except for Pistia stratiotes, Paspalidium germinatum and Polygonum setosulum. There was a higher percentage cover of Hygrophylla auriculata, Ceratophyllum submersum, Cyperus laevigatus, Utricularia inflexa, Nymphaea lotus, Aeschenomene cristata, Typha domingensis and Eichhornia crassipes at the southern sites. Meanwhile Pycreus nitidus and Ludwigia stolonifera occurred only at the sites located within the southern part of the lake. Percentage cover of Aeschenomene pfundii was highest at sites located at the northern part of the lake. Based on the percentage cover data, the Euclidean clustering of the sampling sites based on the vegetation separated the sites as shown in Figure 2. Based on the vegetation species cover recorded, sites C3, S1 and S3 displayed close similarity; as was N1 and N2. Meanwhile sites C2, S2 and N3 were dissimilar with each other and also displayed distance similarity compared to the other sites.

Table 2: Summary of physico-chemical parameters (mean values + SD) in Lake Baringo at different sampling sites during the period between January and December 2015

	Sampling sites								
Variables	S1	S2	S3	C1	C2	C3	N1	N2	N3
Temperature (°C)	25.6 ± 1.4	26.8 ± 1.3	25.7 ± 0.9	24.7 ± 0.7	24.5 ± 1.1	24.8 ± 0.9	24.2 ± 0.8	24.2 ± 1.2	23.5 ± 0.6
pН	7.44 ± 0.07	7.56 ± 0.14	7.65 ± 0.11	7.88 ± 0.12	7.97 ± 0.09	7.95 ± 0.14	8.72 ± 0.11	8.42 ± 0.09	8.57 ± 0.10
Conductivity (µS cm ⁻¹)	601.2 ± 55.6	612.3 ± 29.2	622.2 ± 65.6	889.2 ± 71.2	892.1 ± 60.2	1101.2 ± 50.2	1199.3 ± 90.2	1171.2 ± 88.9	1167.9 ± 89.2
Turbidity (NTU)	47.4 ± 5.1	50.2 ± 6.7	61.2 ± 7.6	30.2 ± 4.5	37.4 ± 4.1	36.8 ± 3.4	24.8 ± 4.5	26.7 ± 5.4	27.8 ± 3.9
TDS (mg/L)	212.3 ± 26.5	198.0 ± 20.9	204.5 ± 21.3	157.0 ± 19.4	189.0 ± 16.7	165.0 ± 20.3	153.0 ± 16.5	155.0 ± 17.8	153.2 ± 18.5
DO (mg/L)	5.52 ± 1.02	5.43 ± 0.98	5.73 ± 0.96	4.62 ± 0.77	4.83 ± 0.86	4.71 ± 0.83	4.63 ± 0.78	4.73 ± 0.93	4.65 ± 0.99
Salinity (‰)	0.46 ± 0.09	0.52 ± 0.12	0.53 ± 0.08	0.58 ± 0.09	0.56 ± 0.08	0.57 ± 0.07	0.56 ± 0.08	0.57 ± 0.07	0.54 ± 0.08
Alkalinity (mg/L)	124.5 ± 13.4	129.2 ± 19.2	134.2 ± 20.3	199.2 ± 23.3	191.2 ± 19.8	190.2 ± 21.2	228.2 ± 18.8	225.2 ± 23.4	234.7 ± 19.2
NO ₃ -N (mg/L)	8.2 ± 1.3	8.0 ± 1.1	8.9 ± 1.3	6.1 ± 1.1	5.6 ± 0.9	5.6 ± 0.8	2.4 ± 0.3	2.5 ± 0.5	2.4 ± 0.3
Dissolved Organic P [DOP] (mg/L)	2.8 ± 0.9	2.6 ± 0.8	2.3 ± 0.7	1.8 ± 0.3	1.7 ± 0.4	1.6 ± 0.3	1.2 ± 0.2	1.3 ± 0.3	1.1 ± 0.2
Dissolved organic carbon [DOC] (mg/L)	6.5 ± 1.5	7.8 ± 1.2	6.6 ± 1.1	8.7 ± 1.8	8.6 ± 1.5	9.1 ± 1.6	8.4 ± 1.1	8.2 ± 1.2	8.6 ± 1.4

Table 3: Macrophyte species composition, habitat forms and percent cover (range) at the southern stations (S1, S2 and S3), central stations (C1, C2 & C3) and northern stations (N1, N2 & N3) of Lake Baringo during the study

Family	Species	Habit/Form	Percent cover ranges at the study stations								
			S1	S2	S3	C1	C2	C3	N1	N2	N3
Acanthaceae	Hygrophylla auriculata (Schum.) Heine	Emerged	1-5	1-5	5-10	ı	1-3	1-5			-
Araceae	Pistia stratiotes L.	Free floating	4-15	11-20	30-50	10-15	10-15	15-20	5-10	5-15	5-15
Asteraceae	Adenostemma caffrum DC.	Emerged	1-3	1	1-5	ı	1	-	1	ı	1-3
Azollaceae	Azollaceae Azolla nilotica Decne ex Mett.		-	5-10	5-10	ı	1	-	1	ı	-
	Azolla pinnata R. Br.	Free floating	-	1	5-10	ı	1	-	1	ı	-
Ceratophyllaceae	Ceratophyllum dermasum L.	Submerged						30-60			
	Ceratophyllum submersum L.	Submerged	50-75	50-75	15-25	30-40	50-75	30-60	20-35	30-50	30-50
Chlorophyceae	<i>Spyrogyra</i> sp	Submerged	-	1	10-20	0-5	5-10	-	1	ı	-
Convolvulaceae	Ipomoea aquatica Forsk	Emerged	10-20	1	10-20	ı	20-30	15-25	10-15	ı	10-15
Cyperaceae	Cyperus articulatus L.	Emerged	-	-	10-20			10-20		-	-
	Cyperus laevigatus Makaloa	Emerged	20-30	20-30	20-30	ı	1	20-30	1	15-20	-
	Cyperus rotundus L.	Emerged	-	15-30	15-30			1-5	1	1-5	-
	Pycreus nitidus (Lam.) J. Raynal.	Emerged			20-30	ı	1	-	1	1-5	-
Lentibulariaceae	Utricularia inflexa Forssk.	Submerged	5-10	1	10-20	ı	1	1-5		1-5	1-5
Nymphaeaceae	Nymphaea lotus L.	Free floating	10-20	20-33	20-30	5-14	11-18	20-25	10-15	1	1
Onagraceae	Ludwigia abyssinica A. Rich	Emerged			1-5	1-5	1-5			10-15	-
	Ludwigia leptocarpa (Nutt.) Hara	Emerged	20-30	-	-	1-3	-	-	-	-	-
	Ludwigia stolonifera (Guill & Perr) P.H. Raven	Emerged/Rooted floating			18-25	-	-	-	-	-	-
Pappilionaceae	Aeschenomene cristata Vatke	Emerged	30-50	30-40	25-40	10-20	10-20	30-45	10-15	20-35	50-75
	Aeschenomene pfundii Taub	Emerged	0-4	20-30	20-30	0-3	1-2	10-15	5-20	25-40	25-60
	Sesbania sesban (L.) Merril	Emergent	16-25	-	-	1-3	-	-	1-5	1-5	1-5
Poaceae	Echinochloa stagnina (Retz.) P. Beauv.	Emergent			22-28	1-5	5-10	-	-	-	-
	Leersia hexandra Sw.	Emerged	20-30	20-30	30-60	-	-	30-40	-	20-30	18-24
	Panicum repens L.	Rooted floating	-	-	-	-	11-14	9-14	10-16	10-15	-
	Paspalidium germinatum (Forsk) Stapf.	Emerged/Rooted floating	30-40	10-24	11-18	-	17-30	-	30-50	12-22	25-50
	Rottboelia exaltata L.f.	Emerged/Rooted floating	0-3	4-8	30-55	-	14-25	3-12	-	30-50	28-52
Polygonaceae	Polygonum salicifolium Willd.	Emerged	2-10	3-12	3-10	-	-	-	1-5	3-9	3-11
	Polygonum setosulum	Emerged	4-9	11-16	4-16	5-18	7-17	5-15	3-11	4-22	13-20
Typhaceae	Typha domingensis Pers.	Emerged	10-30	12-28	12-29	5-25	5-25	5-30	5-9	3-16	5-21
Pontederiaceae	Eichhornia crassipes (Mart.) Solms	Free floating	10-20	20-30	10-20	3-5	7-11	10-20	2-3	2-3	2-3

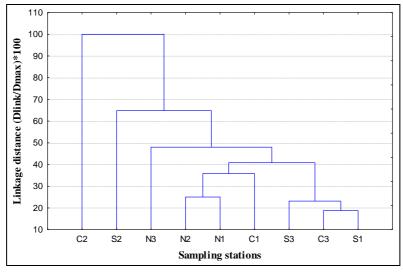


Fig 2: Dendrogram illustrating classifications of stations based on vegetation cover (scaling to dlink/dmax*100) in Lake Baringo during the sampling period.

The relationships among environmental variables and vegetation species cover within the lake are shown in Figure 3. Four principle factors (eigen values > 1) were extracted to explain the variability in the PCA and together, two main factors explained 58.5% of the total data variance. Species such as *Echinochloa stagnina*, *Azolla pinnata*, *Typha domingensis*, *Pistia stratiotes* and *Polygonum setosulum* were positively associated with the DOC, alkalinity and conductivity. Meanwhile the *Nymphaea lotus*, *Ceratophyllum submersum*, *Adenostemma caffrum*, *Ludwigia leptocarpa*,

Rottboelia exaltata, Azolla nilotica, Cyperus laevigatus, Cyperus articulatus, Cyperus rotundus and Paspalidium germinatum were associated with temperature, pH, TDS and DO. Ipomoea aquatica, Spyrogyra sp., Ludwigia abyssinica, Polygonum salicifolium, Sesbania sesban and Utricularia inflexa were associated with salinity, NO₃-N, DOP and turbidity. Hygrophylla auriculata, Eichhornia crassipes, Aeschenomene pfundii and Hygrophylla auriculata were not associated with any environmental variable.

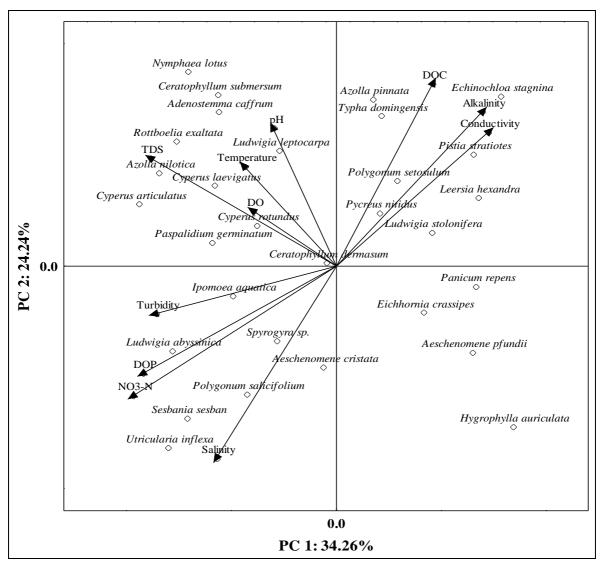


Fig 3: Results of Principal Component Analysis (PCA) on complete environmental variable vectors and species structure in Lake Baringo during the study period.

4. Discussion

The Lake Baringo environment displayed considerable spatial variation, where sampling sites situated near the inlets had higher turbidity, TDS, DO, NO₃-N and DOP but lower pH, conductivity, salinity, alkalinity and DOC. Most of the physical and chemical water quality parameters are similar to studies conducted earlier within the same lake ^[56]. Previous studies in the lake indicate that variation in water quality variables are related to prevailing weather changes driven mainly by the rainfall pattern within the catchment ^[57] as well as presence of hydrothermal recharge in the northern part of the lake ^[46]. The differences in water temperature at different sites therefore were due to the inflow of cooler water into the lake from the catchment. Deeper water in the sites located at

the northern parts of the lake translates to relatively larger water mass which takes longer to warm up and cool down. The relatively high pH at the northern sites is due to high concentrations of carbonate salts, typically sodium carbonate (and related salt complexes), giving rise to high alkalinity. Nevertheless, the high pH in this lake compared to many freshwater lakes could be attributed to lack of any surface outlet. Although there is emerging evidence of underground outlet in this lake [46], such outlets may not be sufficient to prevent accumulation of salts in water. The relatively low pH at the inlet sites could therefore be attributed to inflow of freshwater from the incoming rivers Molo, Ol Arabel, Endau and Perkerra. The same reason could be linked to the lower conductivity, salinity and alkalinity at the inlet sites compared

to more offshore areas because of the inflow of freshwater with low dissolved substances. The high turbidity, TDS and DOP recorded in the lake especially at the inlet sites was attributed to the resuspension of the organic matter in the inflow water mainly from areas experiencing diverse human activities such as logging, charcoal burning, livestock farming and other agricultural activities [47, 48, 58, 59] Therefore the heterogeneity of the environmental variables across the sampling sites within the lake can be characterized by the sites located in the southern zones and apparent lack of outlets for sites in the central and southern parts of the lake.

There are no previous studies reporting on macrophytes in Lake Baringo. In the present study, 16 macrophyte families and 30 aquatic vascular plant species were identified in Lake Baringo with majority of the families being represented by just a single species. This indicate low number of macrophyte families compared to other tropical freshwater lakes [60, 9]. The highest macrophyte species composition was recorded around River Molo mouth and the central zone around Ruko conservancy probably due to rich inflow of nutrients into the lake from the catchment areas. Species belonging to the families Cyperaceae and Poaceae were the dominant groups and with wide distribution in the sampled sites, which has also been reported in other tropical waterbodies [6, 60-66]. Poaceae and Cyperaceae, which are among the bestrepresented families, are also the most important families in other freshwater ecosystems due to their ability to tolerate wide range of environmental conditions [61, 66-69]. Occurrence of several families with few species indicates that the growing conditions are favorable for survival of a few species of macrophytes. This study also established a mixture of macrophyte life forms, with a dominance of emergent forms. Such forms are important in freshwater ecosystems [70-76], suggesting macrophyte adaptation to conditions of the freshwater as opposed to saline conditions. Dominance of emergent species is an expected finding in many tropical lakes, due to nutrients in the water column that can support their growth. Free-floating species were the second most important group in terms of frequency of occurrence. The elevated frequency and richness of this group may be directly associated with the high nutrient status of the lake [77-80].

This first detailed floristic characterization for the Lake Baringo, which is an endorheic freshwater lake, suggest a unique pattern of macrophyte species distribution in terms of percentage cover, with the highest cover for most species being observed near the inlets sites. This suggests that the inflow of freshwater with nutrients from the catchment drives the vegetation community structure of the lake. Any combination of high DOP, NO₃-N, TDS and low DO could explain the high percentage occurrence of macrophytes near the inlet sites. It is important to note that there were four dominant species in terms of percent cover: Ceratophyllum dermasum, Ceratophyllum submersum, Cyperus laevigatus and Eicchornia crassipes, which may be attributed to their prolific multiplication and growth habit in nitrogen and phosphorus rich waters [81]. Moreover, the proliferation of the E. crassipes, a weed that thrives under conditions of contaminated or nutrient rich water [82-84] showed the extent to which water in these locations was nutrient rich. Nonetheless, there is no previous study that has determined how diverse environmental factors drive the macrophyte assemblage in the

The findings of this study on the relationships between aquatic macrophytes and environmental variables established

that four principle factors (eigen values > 1) explained the variability in macrophyte assemblage. Different combinations of environmental variables are necessary to explain different assemblage attributes. Species such as E. stagnina, A. pinnata, T. domingensis, Pi. stratiotes and Po. setosulum were explained by variation in DOC, alkalinity and conductivity as they have been associated to grow better in areas with high carbon content and low pH [85], while Nymphaea lotus, Cer. submersum, A. caffrum, L. leptocarpa, R. exaltata, A. nilotica, Cyp. laevigatus, Cyp. articulatus, Cyp. rotundus and P. germinatum were caused by variation in temperature, pH, TDS and DO. Some of these plants have been established to grow in nutrient deficient waterbodies and therefore appear not dependent on the nutrients for survival [86]. Some species such as I. aquatica, Spyrogyra sp., L. abyssinica, Po. salicifolium, Ses. sesban and U. inflexa were associated with salinity, NO₃-N, DOP and turbidity. Temperature, pH, TDS, DO, alkalinity, conductivity, salinity, NO₃-N, DOP and turbidity are important explanatory variables of macrophyte richness and that combinations of different processes generate different patterns of macrophyte richness in the lacustrine habitats. For example, depth may have been important to explain the macrophyte richness because macrophytes are commonly organized along depth gradients in the littoral zone of water bodies [62]. Indeed, as diversity of plant species community structure changed increases, due complementary interactions with the environmental factors.

5. Conclusion

The present results showed that a simple set of variables was sufficient to explain macrophyte assemblages in a relatively small subtropical lake. Thus, a combination of factors including physical and chemical variables was an important explanation of richness and community structure. A combination of water quality and nutrients, best explained macrophyte richness. The possibility that such species depend exclusively on certain conditions in the lakes, suggests that the sites of the Lake Baringo should be considered important areas for assessing their ecological role in the lake.

6. Acknowledgements

We are grateful to the Government of Kenya for funding the study in Lake Baringo. The lead author benefited from the monthly facilitation by Kenya Marine and Fisheries Research Institute (KMFRI) through KMFRI Baringo management. We acknowledge Bernard Wanjohi of Department of Biological Sciences, University of Eldoret in identifying plant specimens and Alfred Atieno of Department of Fisheries and Aquatic Sciences, University of Eldoret, for summarizing data.

7. References

- 1. Revenga C, Brunner J, Henninger N, Kassem K, Payne R. Pilot Analysis of Global Ecosystems: Freshwater Ecosystems. World Resources Institute (WRI), Washington DC, 2000.
- 2. Chambers PA, Lacoul P, Murphy KJ, Thomaz SM. Global diversity of aquatic macrophytes in freshwater. Hydrobiologia. 2008; 595(1):9-26.
- 3. Denny P. Submerged and floating-leaved aquatic macrophytes (euhydrophytes). In The ecology and management of African wetland vegetation Springer, Dordrecht, 1985, 19-42.
- 4. Henry-Silva GG, Camargo AF, Pezzato MM. Growth of free-floating aquatic macrophytes in different

- concentrations of nutrients. Hydrobiologia, 2008; 610(1):153.
- 5. Wetzel RG. Fundamental processes within natural and constructed wetland ecosystems: short-term versus long-term objectives. Water Science and Technology. 2001; 44(11-12):1-8.
- Tamire G, Mengistou S. Macrophyte species composition, distribution and diversity in relation to some physicochemical factors in the littoral zone of Lake Ziway, Ethiopia. African Journal of Ecology. 2013; 51(1):66-77.
- 7. Biudes JFV, Camargo AFM. Studying the limitant factors to primary production of aquatic macrophytes in Brazil. Oecologia Australis. 2008; 12(1):7-19.
- 8. Thomaz S, Esteves F, Murphy K, Dos Santos A, Caliman A, Guariento R. Aquatic macrophytes in the tropics: ecology of populations and communities, impacts of invasions and human use. International commission on tropical biology and natural resources. Tropical biology and conservation management. 2011; 4:27-60.
- 9. Bando FM, Michelan TS, Cunha ER, Figueiredo BR, Thomaz SM. Macrophyte species richness and composition are correlated with canopy openness and water depth in tropical floodplain lakes. Brazilian Journal of Botany. 2015; 38(2):289-294.
- Barko JW, James WF. Effects of submerged aquatic macrophytes on nutrient dynamics, sedimentation, and resuspension. In The structuring role of submerged macrophytes in lakes Springer, New York, NY, 1998, 197-214.
- 11. Greenway M. The role of macrophytes in nutrient removal using constructed wetlands. In Environmental bioremediation technologies. Springer, Berlin, Heidelberg, 2007, 331-351.
- 12. Scheffer M, Jeppesen E. Regime shifts in shallow lakes. Ecosystems. 2007; 10(1):1-3.
- 13. Lacoul P, Freedman B. Environmental influences on aquatic plants in freshwater ecosystems. Environmental Reviews. 2006; 14(2):89-136.
- 14. Demars BO, Harper DM. Distribution of aquatic vascular plants in lowland rivers: separating the effects of local environmental conditions, longitudinal connectivity and river basin isolation. Freshwater Biology. 2005; 50(3):418-437.
- 15. Kohler A, Schneider S. Macrophytes as bioindicators. *Large Rivers*. 2003; 14(1, 2):17-31.
- Melzer A. Aquatic macrophytes as tools for lake management. In The Ecological Bases for Lake and Reservoir Management Springer, Dordrecht, 1999, 181-190
- 17. Williams P, Whitfield M, Biggs J, Bray S, Fox G, Nicolet P *et al.* Comparative biodiversity of rivers, streams, ditches and ponds in an agricultural landscape in Southern England. Biological conservation. 2004; 115(2):329-341.
- 18. McElarney YR, Rippey B. A comparison of lake classifications based on aquatic macrophytes and physical and chemical water body descriptors. Hydrobiologia. 2009; 625(1):195.
- 19. Roberto MC, Santana NF, Thomaz SM. Limnology in the Upper Paraná River floodplain: large-scale spatial and temporal patterns, and the influence of reservoirs. Brazilian Journal of Biology. 2009; 69(2):717-725.
- 20. Riis T, Biggs BJ. Hydrologic and hydraulic control of

- macrophyte establishment and performance in streams. Limnology and oceanography. 2003; 48(4):1488-1497.
- 21. Lacoul P, Freedman B. Relationships between aquatic plants and environmental factors along a steep Himalayan altitudinal gradient. Aquatic Botany. 2006; 84(1):3-16.
- 22. Chappuis E, Gacia E, Ballesteros E. Environmental factors explaining the distribution and diversity of vascular aquatic macrophytes in a highly heterogeneous Mediterranean region. Aquatic Botany. 2014; 113:72-82.
- 23. Schneider B, Cunha ER, Marchese M, Thomaz SM. Explanatory variables associated with diversity and composition of aquatic macrophytes in a large subtropical river floodplain. Aquatic Botany. 2015; 121:67-75.
- 24. Barko JW, Adams MS, Clesceri NL. Environmental factors and their consideration in the management of submersed aquatic vegetation: a review. Journal of Aquatic Plant Management. 1986; 24(1):1-10.
- 25. Heegaard E, Birks HH, Gibson CE, Smith SJ, Wolfe-Murphy S. Species—environmental relationships of aquatic macrophytes in Northern Ireland. Aquatic botany. 2001; 70(3):175-223.
- 26. Gacia E, Ballesteros E, Camarero L, Delgado O, Palau A, Riera JL *et al.* Macrophytes from lakes in the eastern Pyrenees: community composition and ordination in relation to environmental factors. Freshwater Biology. 1994; *32*(1):73-81.
- 27. Scheffer M, van Nes EH. Shallow lakes theory revisited: various alternative regimes driven by climate, nutrients, depth and lake size. Hydrobiologia. 2007; 584(1):455-466.
- 28. Vestergaard O, Sand-Jensen K. Alkalinity and trophic state regulate aquatic plant distribution in Danish lakes. Aquatic Botany. 2000; 67(2):85-107.
- 29. Palmer MA, Covich AP, Lake SAM, Biro P, Brooks JJ, Cole J *et al.* Linkages between Aquatic Sediment Biota and Life Above Sediments as Potential Drivers of Biodiversity and Ecological Processes: A disruption or intensification of the direct and indirect chemical, physical, or biological interactions between aquatic sediment biota and biota living above the sediments may accelerate biodiversity loss and contribute to the degradation of aquatic and riparian habitats. AIBS Bulletin. 2000; 50(12):1062-1075.
- 30. Feldmann T, Noges P. Factors controlling macrophyte distribution in large shallow Lake Võrtsjärv. Aquatic Botany. 2007; 87(1):15-21.
- 31. Hayashi M, Rosenberry DO. Effects of ground water exchange on the hydrology and ecology of surface water. Groundwater. 2002; 40(3):309-316.
- 32. Bornette G, Puijalon S. Response of aquatic plants to abiotic factors: a review. Aquatic Sciences, 2011; 73(1):1-14.
- 33. Chen X, Li X, Xie Y, Li F, Hou Z, Zeng J. Combined influence of hydrological gradient and edaphic factors on the distribution of macrophyte communities in Dongting Lake wetlands, China. Wetlands ecology and management. 2015; 23(3):481-490.
- 34. Van Geest GJ, Coops H, Roijackers RMM, Buijse AD, Scheffer M. Succession of aquatic vegetation driven by reduced water-level fluctuations in floodplain lakes. Journal of Applied Ecology. 2005; 42(2):251-260.
- 35. Radomski P, Goeman TJ. Consequences of Human Lakeshore Development on Emergent and Floating-Leaf Vegetation Abundance. North American Journal of

- Fisheries Management. 2001; 21(1):46-61.
- 36. Alexander ML, Woodford MP, Hotchkiss SC. Freshwater macrophyte communities in lakes of variable landscape position and development in northern Wisconsin, USA. Aquatic Botany. 2008; 88(1):77-86.
- 37. Hicks AL, Frost PC. Shifts in aquatic macrophyte abundance and community composition in cottage developed lakes of the Canadian Shield. Aquatic Botany, 2011; 94(1):9-16.
- 38. Nyström P, Strand J. Grazing by a native and an exotic crayfish on aquatic macrophytes. Freshwater Biology, 1996; 36(3):673-682.
- 39. Rørslett B. Environmental factors and aquatic macrophyte response in regulated lakes-a statistical approach. Aquatic Botany. 1984; 19(3, 4):199-220.
- 40. Rosenberg DM, McCully P, Pringle CM. Global-scale environmental effects of hydrological alterations: introduction. Bio Science. 2000; 50(9):746-751.
- 41. Murphy KJ. Plant communities and plant diversity in softwater lakes of northern Europe. Aquatic botany. 2002; 73(4):287-324.
- 42. Alahuhta J, Heino J. Spatial extent, regional specificity and metacommunity structuring in lake macrophytes. Journal of Biogeography. 2013; 40(8):1572-1582.
- 43. Sand-Jensen K, Riis T, Vestergaard O, Larsen SE. Macrophyte decline in Danish lakes and streams over the past 100 years. Journal of Ecology. 2000; 88(6):1030-1040.
- 44. Dodson SI, Lillie RA, Will-Wolf S. Land use, water chemistry, aquatic vegetation, and zooplankton community structure of shallow lakes. Ecological Applications. 2005; 15(4):1191-1198.
- 45. Havens KE. Submerged aquatic vegetation correlations with depth and light attenuating materials in a shallow subtropical lake. Hydrobiologia. 2003; 493(1-3):173-186.
- 46. Tarits C, Renaut RW, Tiercelin JJ, Le Hérissé A, Cotten J, Cabon JY. Geochemical evidence of hydrothermal recharge in Lake Baringo, central Kenya Rift Valley. Hydrological processes. 2006; 20(9):2027-2055.
- 47. Aloo PA. Effects of climate and human activities on the ecosystem of Lake Baringo, Kenya. In The East African great lakes: Limnology, palaeolimnology and biodiversity. Springer, Dordrecht, 2002, 335-347.
- 48. Kiage LM, Liu KB, Walker ND, Lam N, Huh OK. Recent land-cover/use change associated with land degradation in the Lake Baringo catchment, Kenya, East Africa: evidence from Landsat TM and ETM+. International Journal of Remote Sensing. 2007; 28(19):4285-4309.
- 49. Owen RB, Renaut RW, Hover VC, Ashley GM, Muasya AM. Swamps, springs and diatoms: wetlands of the semi-arid Bogoria-Baringo Rift, Kenya. Hydrobiologia. 2004; 518(1-3):59-78.
- 50. Kallqvist T. Primary production and phytoplankton in Lake Baringo and Naivasha, Kenya. Norwegian Institute for Water Research Report, Blinden, Oslo, 1987, 59.
- Cook CD, Gut BJ, Rix EM, Schneller J. Water plants of the world: a manual for the identification of the genera of freshwater macrophytes. Springer Science & Business Media, 1974.
- 52. Federation WE, Association APH. Standard methods for the examination of water and wastewater. American Public Health Association (APHA): Washington, DC, USA, 2005.

- 53. Legendre P, Gallagher ED. Ecologically meaningful transformations for ordination of species data. Oecologia. 2001; 129(2):271-280.
- 54. Lasaponara R. On the use of principal component analysis (PCA) for evaluating interannual vegetation anomalies from SPOT/VEGETATION NDVI temporal series. Ecological Modelling. 2006; 194(4):429-434.
- 55. Brauner N, Shacham M. Role of range and precision of the independent variable in regression of data. AIChE journal. 1998; 44(3):603-611.
- 56. Ouma H, Mwamburi J. Spatial variations in nutrients and other physicochemical variables in the topographically closed Lake Baringo freshwater basin (Kenya). Lakes & Reservoirs: Research & Management. 2014; 19(1):11-23.
- 57. Omondi R, Kembenya E, Nyamweya C, Ouma H, Machua SK, Ogari Z. Recent limnological changes and their implication on fisheries in Lake Baringo, Kenya. Journal of Ecology and The Natural Environment. 2014; 6(5):154-163.
- 58. Odada EO, Onyando JO, Obudho PA. Lake Baringo: Addressing threatened biodiversity and livelihoods. Lakes & Reservoirs: Research & Management. 2006; 11(4):287-299.
- 59. Hickley P, Muchiri M, Boar R, Britton R, Adams C, Gichuru N *et al.* Habitat degradation and subsequent fishery collapse in Lakes Naivasha and Baringo, Kenya. International Journal of Ecohydrology & Hydrobiology. 2004; 4(4):503-517.
- 60. Adesina GO, Akinyemiju OA, Muoghalu JI. Checklist of the aquatic macrophytes of Jebba Lake, Nigeria. Ife Journal of Science. 2011; 13(1):93-102.
- 61. Chambers PA, Lacoul P, Murphy KJ, Thomaz SM. Global diversity of aquatic macrophytes in freshwater. Hydrobiologia. 2008; 595(1):9-26.
- 62. Dos Santos AM, Thomaz S. Aquatic macrophytes diversity in lagoons of a tropical floodplain: the role of connectivity and water level. Austral Ecology. 2007; 32(2):177-190.
- 63. Ekblom A, Stabell B. Paleohydrology of Lake Nhaucati (southern Mozambique),~ 400 AD to present. Journal of Paleolimnology. 2008; 40(4):1127.
- 64. Ficken KJ, Wooller MJ, Swain DL, Street-Perrott FA, Eglinton G. Reconstruction of a subalpine grass-dominated ecosystem, Lake Rutundu, Mount Kenya: a novel multi-proxy approach. Palaeogeography, Palaeoclimatology, Palaeoecology. 2002; 177(1-2):137-149
- 65. Kennedy MP, Lang P, Grimaldo JT, Martins SV, Bruce A, Lowe S *et al.* The Zambian Macrophyte Trophic Ranking scheme, ZMTR: A new biomonitoring protocol to assess the trophic status of tropical southern African rivers. Aquatic Botany. 2016; 131:15-27.
- 66. Mackay SJ, James CS, Arthington AH. Macrophytes as indicators of stream condition in the wet tropics region, Northern Queensland, Australia. Ecological Indicators. 2010; 10(2):330-340.
- 67. Capon SJ, Brock MA. Flooding, soil seed bank dynamics and vegetation resilience of a hydrologically variable desert floodplain. Freshwater Biology. 2006; 51(2):206-223
- 68. Stutz S, Borel CM, Fontana SL, Del Puerto L, Inda H, García-Rodriguez F *et al.* Late Holocene climate and environment of the SE Pampa grasslands, Argentina, inferred from biological indicators in shallow, freshwater

- Lake Nahuel Rucá. Journal of Paleolimnology. 2010; 4(3):761-775.
- 69. Michelan TS, Thomaz S, Mormul RP, Carvalho P. Effects of an exotic invasive macrophyte (tropical signalgrass) on native plant community composition, species richness and functional diversity. Freshwater Biology. 2010; 55(6):1315-1326.
- 70. Kiss MK, Lakatos G, Borics G, Gidó Z, Deák C. Littoral macrophyte–periphyton complexes in two Hungarian shallow waters. Hydrobiologia. 2003; 506(1-3):541-548.
- 71. Nurminen L. Macrophyte species composition reflecting water quality changes in adjacent water bodies of lake Hiidenvesi, SW Finland. In Annales Botanici Fennici Finnish Zoological and Botanical Publishing Board, 2003, 199-208.
- 72. Qiu D, Wu Z, Liu B, Deng J, Fu G, He F. The restoration of aquatic macrophytes for improving water quality in a hypertrophic shallow lake in Hubei Province, China. Ecological Engineering. 2001; 18(2):147-156.
- 73. Cazzanelli M, Warming TP, Christoffersen KS. Emergent and floating-leaved macrophytes as refuge for zooplankton in a eutrophic temperate lake without submerged vegetation. Hydrobiologia. 2008; 605(1):113-122.
- 74. Vis C, Hudon C, Carignan R. An evaluation of approaches used to determine the distribution and biomass of emergent and submerged aquatic macrophytes over large spatial scales. Aquatic Botany. 2003; 77(3):187-201.
- 75. Hough RA, Fornwall MD, Negele BJ, Thompson RL, Putt DA. Plant community dynamics in a chain of lakes: principal factors in the decline of rooted macrophytes with eutrophication. Hydrobiologia. 1989; 173(3):199-217.
- 76. Vyas V, Yousuf S, Bharose S, Kumar A. Distribution of macrophytes in River Narmada near water intake point. Journal of Natural Sciences Research. 2012; 2(3):54-59.
- 77. Meerhoff M, Mazzeo N, Moss B, Rodríguez-Gallego L. The structuring role of free-floating versus submerged plants in a subtropical shallow lake. Aquatic Ecology. 2003; *37*(4):377-391.
- 78. Netten JJ, Van Zuidam J, Kosten S, Peeters ET. Differential response to climatic variation of free-floating and submerged macrophytes in ditches. Freshwater Biology. 2011; 56(9):1761-1768.
- 79. O'Farrell I, Izaguirre I, Chaparro G, Unrein F, Sinistro R, Pizarro H *et al.* Water level as the main driver of the alternation between a free-floating plant and a phytoplankton dominated state: a long-term study in a floodplain lake. Aquatic Sciences. 2011; 73(2):275-287.
- 80. Henry-Silva GG, Camargo AF, Pezzato MM. Growth of free-floating aquatic macrophytes in different concentrations of nutrients. Hydrobiologia, 2008; 610(1):153.
- 81. Gichuki J, Omondi R, Boera P, Okorut T, Matano AS, Jembe T *et al.*, Water Hyacinth Eichhornia crassipes (Mart.) Solms-Laubach dynamics and succession in the Nyanza Gulf of Lake Victoria (East Africa): implications for water quality and biodiversity conservation. The Scientific World Journal, 2012.
- 82. Williams AE, Duthie HC, Hecky RE. Water hyacinth in Lake Victoria: Why did it vanish so quickly and will it return?. Aquatic Botany. 2005; 81(4):300-314.
- 83. Williams AE, Hecky RE. Invasive aquatic weeds and

- eutrophication: The case of water hyacinth in Lake Victoria. Restoration and management of tropical eutrophic lakes. *5*, 0-00, 2005.
- 84. Moyo P, Chapungu L, Mudzengi B. Effectiveness of water Hyacinth (*Eichhornia crassipes*) in remediating polluted water: The case of Shagashe River in Masvingo, Zimbabwe. ecosystems, 2013; 7:9.
- 85. Wong SC, Osmond CB. Elevated atmosphere partial pressure of CO2 and plant growth. III. Interactions between *Triticum aestivum* (C3) and *Echinochloa frumentacea* (C4) during growth in mixed culture under different CO2, N nutrition and irradiance treatments, with emphasis on below-ground responses estimated using the δ13C value of root biomass. Functional Plant Biology. 1991; 18(2):137-152.
- 86. Kateyo EM. Biodiversity of an interface zone of a nutrient-deficient lake (Nabugabo) in Uganda: macrophytes. African journal of ecology. 2007; 45(2):130-134.