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Metals, Petroleum Hydrocarbons and Organochlorines in Inshore Sediments and Waters of Mombasa, Kenya

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A coastal zone pollution monitoring programme for developing countries, the Land-Ocean Contamination Study (LOCS), was initiated by the British Geological Survey (BGS) under funding from the United Kingdom Overseas Development Administration in 1995. The central objective of LOCS is the provision of contaminant monitoring, impact amelioration and integrated coastal zone management protocols to meet the specific social, technical and economic requirements of tropical developing regions.

A geochemical and oceanographic survey of the inshore waters and sediments of Mombasa, Kenya was undertaken by BGS in liaison with the Kenya Marine and Fisheries Research Institute as a component of the LOCS programme during the period September 1995–January 1996 (Williams *et al.*, 1996a). The administrative district of Mombasa occupies a 275-km² area on the Indian Ocean coast, centred on latitude 4.05°S and longitude 39.65°E (Fig. 1). The

district comprises the island settlement of Mombasa (now connected to the mainland by the *ca* 1 km Makupa causeway) and a series of urban developments to the north-east, west and south-west, the combined population of which exceeds 500 000. Mombasa Island is bounded by the Port Kilindini and Port Reitz creek complexes to the south and west and by the Mombasa Harbour–Port Tudor complex to the east and north. Fluvial inputs to these systems are provided by the Shimba, Mwache, Komboni and Tsalu Rivers, each with annual discharges in the range $10-50 \times 10^6 \text{m}^3$. The open coastline to the north-east and south-west of Mombasa is fringed by a multiple reef (typically 1– 1.5 km offshore) which extends virtually unbroken for a distance of 60 km between Mtwapa and Gazi.

The vulnerability of both the inshore and nearshore environments to contamination arising from the urban and industrial expansion of Mombasa (compounded by recent tourist development) has been widely documented (e.g. UNEP, 1982). An inventory of industrial activities in Mombasa Municipality (Munga *et al.*, 1994) indicates that suspended solid discharges may currently exceed 21 000 t yr⁻¹. The impact and fate of this contaminant flux within the coastal environment has not, however, previously been ascertained.

In the LOCS survey, samples of water, suspended particulate matter (SPM) and sediment were collected from 48 stations within Mombasa's estuarine creeks and along the reef front extending north-eastwards (in accordance with the prevailing current regime) some 13 km to Mtwapa (Fig. 1). All sites were positioned using a SilvaTM non-differential global positioning system (GPS). Mid-column water samples were collected under both ebb and flood tide conditions during independent campaigns in September 1995 and January 1996. In situ measurements of water temperature, dissolved oxygen (DO), pH, Eh, turbidity, conductivity



Fig. 1 Regional location and physiography of Mombasa, showing the position of LOCS sampling stations.

and salinity were made using a pHOXTM 902 sonde and datalogger. Calibration of the component probes in the sonde was carried out daily using *1*. 2% Na₂SO₃ (zero DO), *2*. pH 4.0 and pH 10.0 buffer solutions, *3*. potassium ferro- and ferricyanide redox standards of 125 and 350 mV, *4*. KCl conductivity standards of 100, 1000 and 50 000 μ S, and *5*. polymeric turbidity solutions of 1, 100 and 1000 FTU. Water samples for trace metal analysis were recovered in a 2.51 acrylic Kemmerer sampler, activated by a messenger. A 0.5-1 volume of sample was immediately vacuum-pumped through a 0.45- μ m cellulose acetate membrane (SartoriusTM) into a HNO₃-washed HDPE bottle (NalgeneTM) and acidified with 1% v/v HNO₃ (ARISTAR). Analysis for a suite of 6 heavy metals (Cd, Cr, Cu, Ni, Pb, Zn)

was carried out by anodic stripping voltammetry (ASV) at the UK Environment Agency laboratory, Llanelli, Wales, to operational detection limits of 0.042 μ g l⁻¹ Cd, 0.35 μ g l⁻¹ Cr, 0.005 μ g l⁻¹ Cu, 0.058 μ g l⁻¹ Ni, 0.024 μ g l⁻¹ Pb and 0.1 μ g l⁻¹ Zn.

Samples of SPM were obtained from a 1-l volume of mid-column water during the filtration of samples for dissolved metal analysis (1995 campaign only). Preweighed cellulose acetate membranes were utilized, thus allowing precise determination of the mass of SPM residues without removal from the filter. Digestion of both the residues and the retaining membranes was carried out at 105° C in sealed PTFE bombs containing 5 ml HNO₃+2 ml HClO₄+2 ml HF. All digests were analysed for 13 trace elements by inductively coupled plasma mass spectrometry (ICP-MS). A suite of pristine filter membranes was prepared and analysed using identical reagents, and appropriate background corrections were applied to the SPM dataset.

Stations characterized by silt and mud sediment lithologies were sampled using a 1.1-m long, 50 mm internal diameter pneumatic piston corer, facilitating the recovery of stratified samples with negligible disturbance of the interfacial layer. Sites with sand or gravel dominated surface lithofacies (in the lower reaches of Mombasa Harbour, Port Kilindini and along the Nyali-Mtwapa reef front) were sampled with a stainless steel Van Veen grab sampler. All cores were retained upright during transportation and were subsampled at 2-5 cm resolution within 24 h of collection. Near-surface redox conditions were established during core extrusion with a Pt electrode and a Hanna InstsTM mV meter. Sediment granulometry was established by wet sieving of core/grab sub-samples to 63 µm, with additional characterisation of the $< 63 \mu m$ fraction using Sedigraph.

The concentrations of major oxides and 13 trace elements in the <150 µm fraction of 252 sediment subsamples (representing a maximum of 20 stratigraphic levels per coring station) were determined by X-ray fluorescence (XRF). The solid phase speciation of 11 metals (Fe, Mn, Al, Co, Cr, Cu, Pb, Zn, V, Ni, Cd) was established for selected sites using an extractive procedure based on that of Breward and Peachey (1983), entailing 1. agitation of 1.g fresh sediment in 1N ammonium acetate solution to remove exchangeable metals, 2. application of 1N ammonia solution to extract the organic fraction, with subsequent isolation of fulvic and humic components by HCl precipitation, 3. partial digestion of the inorganic residue in 0.1N hydroxylamine hydrochloride in 25% v/v acetic acid to extract hydrous Fe and Mn oxides, and 4. total digestion of the inorganic residue in concentrated nitric + perchloric + hydrofluoric acids to liberate metals held in detrital silicates and other residual minerals. All leachates were analysed by inductively coupled plasma emission spectrometry (ICP-ES). Reagent blanks were prepared and simultaneously analysed with each leachate. Surficial sediment from 6 stations, plus sub-surface samples from a Makupa Creek site (MB23) were analysed for n-alkanes, polycyclic aromatic hydrocarbons (PAHs) and organochlorines by a combination of high pressure liquid chromatography (HPLC), gas chromatography (GC) and GC mass-spectrometry (GCMS) following hexane solvent extraction.

Physicochemical data for the 1995 Mombasa seawater suite are given in Table 1. The 1996 suite is not presented, as the temporal variance (based on ANOVA) was found to be < 10%. With the exception of Pb (for which relatively high values of up to $1.9 \ \mu g \ l^{-1}$ were recorded) the median concentrations of dissolved metals Marine Pollution Bulletin

(Cd <0.042 μ g l⁻¹, Cr <0.35 μ g l⁻¹, Cu 2.63 μ g l⁻¹, Ni 0.28 μ g l⁻¹ and Zn 8.1 μ g l⁻¹) were found to lie within the global background range for marine waters (e.g. Fergusson, 1990). However, ANOVA studies of data for six discrete sectors of the study area (A, upper Tudor Creek; B, Port Tudor and Mombasa Harbour; C, Port Kilindini; D, Makupa Creek; E, Port Reitz; and F, the North Coast reef front) highlighted significant local variability. Elevated concentrations of Cu (to 20.5 μ g l^{-1}) and Ni (to 0.65 μ g l^{-1}) in the uppermost reaches of Tudor Creek reflected the natural hydrochemical signature of terrestrial runoff, while a substantial anthropogenic Ni flux into Makupa Creek (probably derived from the adjacent Kibarani landfill site) was indicated by ambient concentrations of Ni up to 0.975 μ g 1⁻¹ in a system characterized by a short (24 h) water residence time (Rees et al., 1996). The reef front waters north of Mombasa (MB41-MB44) showed marked enrichment of Cd, Pb and Zn during both the 1995 and 1996 campaigns, although the derivation of a plausible metal source remains problematic. The recorded Zn concentration range of 24-69 μ g l⁻¹ in these waters exceeded the median for the total Mombasa suite by at least a factor of 3.

Metal concentration data for the Mombasa SPM suite are given in Table 2. Median abundances were found to decline in the order Mn > Zn > Cu >Ni > V > Pb > Cr > As > Co > Th > Mo > U > Cd. Firstrow transition elements (e.g. Mn, V, Cr, Co, Ni), As and Mo showed systematic enrichment in the upper sector of Tudor Creek, relative to all other inshore waters. Zn, although enriched in Tudor Creek, displayed highest concentrations (>6000 mg kg⁻¹ dry weight, with coincident Pb enrichment to 210 mg kg⁻¹) along the North Coast reef front. The lowest concentrations of As, Cd, Cu, Mo, Ni, Pb and Zn were recorded in SPM from Port Reitz. The prevalence of As, Co, Cu, Cr, Ni and V in Port Kilindini and Port Reitz (which host all major shipping, petrochemicals and sewage installations) at concentrations up to an order of magnitude lower than in the largely unindustrialized Tudor Creek catchment suggests that the anthropogenic influence on SPM chemistry is negligible. A more critical control appears to be the contribution of mangrove-derived clastic detritus to the total SPM assemblage. Williams et al. (1996a) have shown this material to accumulate a range of trace cations (both in the mangrove environment and following outwash into the water column) through surface adsorption to virtually ubiquitous Fe/Mn oxide coatings.

Major oxide and trace element data for surficial sediments (the 0-2 cm horizons of piston cores, plus all grab samples) are given in Table 3. The spatial distributions of all major oxides and several minor/ trace transition metals (including Co, Cr, Ni and V) were found to closely reflect sediment provenance and lithology, with no clear anthropogenic overprint. Considerable independence from any lithological con-

	ïZ	0.179	0.595	0.610	0.565	0.512	0.653	0.409	0.327	0.169	0.426	0.414	0.292		0.058	0000	0.280	0.207	0.275	0.325	0.252	0.975	0.858	0.156	0.190	0.058	0.395	0.286	0.277		0.257	700.0	0 287	0110	0.228	0 326	0.124	0.214	0.149	0.260	0.441	0.464	0.282 0.282
-1).	C	< 0.350	< 0.350	< 0.380	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	0.680	0.350	0.470	< 0.350		<0350>		< 0.350	< 0.350	0.450	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	< 0.350	0.410		065.0 /	0300/	0000 ~	< 0.250	< 0.350	< 0.350	0.420	< 0.350	0.350	< 0.350	< 0.350	< 0.350	UCC.N ~
ed in (μg l	Zn	8.120	12.600	9.570	7.800	10.000	12.600	7.740	8.120	18.000	6.780	18.900	8.620		2 000	000-7	9.900	15.600	10.700	9.930	6.780	10.200	7.600	5.220	4.100	2.000	9.120	5.110	6.050		080.0	007.61	4.720	3 100	9.520	67,100	69.600	24.000	49.900	7.800	5.030	11.400	2.370 8.370
tre express	Cu	20.500	6.210	5.850	2.640	2.380	3.040	2.760	2.810	3.880	3.170	2.860	3.110		0.051	100.0	3.170	4.270	1.840	1.670	1.990	2.740	2.200	1.940	2.020	0.051	1.910	2.190	1.640		0707		3 050	A 480	2.210	1 680	3.680	2.880	2.640	1.760	1.260	1.470	100.0
All trace metal data are	Cd	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	> 0.042	< 0.042	< 0.042		< 0.040	7-0-0	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	< 0.042	0.049	< 0.042	< 0.042		< 0.042		< 0.042	< 0.042	< 0.042	0.576	0.676	0.173	0.540	< 0.042	< 0.042	< 0.042	22.640
	Pb	0.571	0.472	0.460	0.460	0.658	0.409	0.536	1.090	1.880	0.923	0.599	0.534		0.074	170.0	0.888	0.830	0.444	0.463	0.950	0.806	0.698	1.970	0.677	0.127	1.290	0.594	1.120		0./94	C7070	0.454	0.648	0.517	1,340	1.080	1.290	1.420	1.400	1.230	0.728	0.677
stermined.	Turb (FTU)	30	44	49	60	39	50	48	52	92	71	67	27	96	69		41	41	44	44	45	54	74	. 09	74		69	11	84	76	5 E		C 2	102	78	51	55	50	51	56	51	49	
were not de	Salinity (%)	33.6	33.4	33.2	33.1	34.4	34.3	34.4	34.6	34.8	34.8	34.7	34.5	34.1	24.2		35.2	35.3	35.3	354	352	350	352	354	353		354	353	352	353			7.00	250	353	354	355	355	355	351	354	354	
ABLE 1 .e that parameters w	nd (μS)	51200	51000	50700	50500	52400	52200	52200	52000	52900	53000	52900	52400	51900	00170		53400	53500	53500	53500	53500	53000	53400	56600	53600		53500	53500	53400	53500	00000	00+00	00400 22200	00700	53500	53700	53500	53800	53900	53300	53500	53500	
	CC b				_										_															_								.,					
T s indicat	Tem] (°C)	27.4	27.4	27.6	27.9	27.5	27.6	27.4	27.0	26.5	26.8	26.9	27.0	27.8	0.12		26.4	26.2	26.7	26.5	26.9	27.5	28.1	27.1	26.8		26.6	26.9	27.4	26.9	8.07 7		4.17	2 L C	27.5	26.4	25.7	25.9	25.7	28.4	26.9	26.4	
ıt data field	DO (% sat)	63.0	63.0	54.0	70.0	80.0	74.0	78.0	0.67	76.0	78.0	75.0	68.0	71.0	0.17		86.0	85.0	0.06	92.0	92.0	65.0	80.0	80.0	88.0		85.0	86.0	85.0	85.0 22 2	070 070	00.0	0.00	60.0	81.0	116.0	98.0	105.0	96.0	102.0	105.0	92.0	
ples. Absen	DO (mg l ⁻¹)	3.4	3.3	3.4	3.7	4.2	3.8	4.0	4.1	3.9	4.1	3.9	3.6	9.6 7	0.0		4.5	4.5	4.6	4.9	4.7	3.3	4.1	4.1	4.6		4.4	4.4	4.4	4.4	7.4	+ + + +	4.4 4.4	, r	4	61	5.0	5.6	5.1	5.2	5.4	4.8	
ombasa seawater samp	Eh(mV)					190	196	190	173								126	85	6	86	,	120	67	93			90	68	67		38	04											
	Hd	7.70	7.60	7.50	7.50	8.00	7.80	7.90	7.90	8.00	8.00	7.90	7.80	7.60	/./0		8.20	8.20	8.20	8.20	8.20	00.8	8.10	8.10	8.10		8.10	8.10	8.00	8.10	8.10 8.00	0.0	06.7	00.0	96	30	8.20	3.30	3.20	8.10	8.20	3.20	
t for Moi	t(°)	.002	. 988	686 [.]	. 186	.028	.017	.025	. 910	.044	.038	.029	.023	900	/00.		.065	010	075		076	035	035	040	.053		.062	.048	.041	040	440. 440.		040	210	020	058	007 2007	86.	361 8	.053 8	.064	.053	
ical data	°) La	4-	+	, L	1.3	14	+ - 4	94	4-	8-4-	4 4	8 -4	1				4-	4	4		-4	- 4	1	4	4-		4 8	- 4	4	4-	+ -	- T	1	ŕ -		4	- 	. – 3	ц. Г.	1 -4	4 - 4	1	
cochemi	Long (39.653	39.644	39.626	39.612	39.651	39.644	39.645	39.652	39.678	39.671	39.673	39.664	39.657	39.040		39.656	39.662	39.664		39.645	39.642	39 63	39.639	39.640		39.643	39.645	39.630	39.630	39.624		20.505	202.02	102.20	C17.95	39.741	39.750	39.765	39.677	39.682	39.681	
Physi	Sample	MB1	MB2	MB3	MB4	MB5	MB6	MB7	MB8	MB9	MB10	MB11	MB12	MB13	MB14	CIGW	MB16	MB17	MB18	MB19	MB21	MB22	MR73	MB24	MB25	MB26	MB27	MB28	MB29	MB30	MB33	PCGIN	ND26		MB30	MR41	MB42	MB43	MB44	MB45	MB46	MB47	M1 1 548
	Reference	E386793	E386794	E386795	E386796	E366797	E386798	E386799	E386600	E386801	E366802	E386803	E386604		E386805	00000077	E386806	E386807	E386808	E386609	E386810	E366811	F386812	E386813	E386814	E386815	E386616	E386817	E386818		E366819	E30062U	E380821	2200022	E300023 F386874	E386875	E386826	E386827	E386828	E386829	E386830	E386831	E380832 MEDIAN

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Concentrations of selected trace elements (mg kg ⁻¹ dry weight) in suspended particulate matter (SPM) from the inshore waters of Mombasa Sample As Cd Cr Co Cu Ph Mr No Ni Th U V Z														
Sample	As	Cd	Cr	Co	Cu	Pb	Mn	Мо	Ni	Th	U	v	Zn	
MB1	11.5	0.50	78.3	14.1	221.1	48.0	516.1	3.40	135.0	9.1	2.6	123.0	303.0	
MB2	12.2	0.40	105.0	18.4	189.5	53.0	690.0	6.00	160.2	10.9	3.0	163.6	359.0	
MB3	10.4	0.67	105.3	19.6	93.6	44.0	745.7	3.80	215.5	10.4	3.0	160.4	412.0	
MB4	16.3	2.20	146.0	27.0	141.5	55.0	113.0	4.20	78.0	14.5	3.8	227.0	442.0	
MB5	7.2	7.90	45.3	8.5	96.1	52.0	423.0	2.40	50.0	4.6	1.5	65.3	412.0	
MB6	14.8	0.42	86.0	11.4	75.0	39.0	528.0	3.30	75.0	7.8	2.5	114.0	238.0	
MB7	15.1	1.10	82.5	14.4	143.1	62.0	628.1	4.00	62.5	9.3	2.8	131.8	435.0	
MB8	16.7	1.70	78.0	13.8	84.2	38.0	560.7	3.90	102.1	8.1	2.6	118.2	467.0	
MB9	5.4	0.35	22.8	4.6	55.7	23.0	229.2	1.50	50.2	3.5	1.1	36.8	272.0	
MB10	6.8	0.16	60.0 -	7.6	96.3	27.0	349.0	3.00	67.1	5.1	1.7	63.3	264.0	
MB11	11.4	0.40	64.0	11.2	137.9	41.0	531.0	3.10	49.2	7.1	2.3	91.3	378.0	
MB12	8.8	1.04	49.1	8.8	104.5	37.0	394.1	2.90	205.0	5.0	1.7	75.4	284.0	
MB16	8.6	0.58	35.1	6.6	95.8	41.0	403.1	3.10	74.6	5.1	2.1	55.6	430.0	
MB17	4.3	0.58	26.3	5.1	60.0	46.0	330.7	2.30	37.3	4.3	1.4	35.8	273.0	
MB18	6.8	0.47	57.2	6.1	113.1	32.0	399.5	2.40	55.2	7.5	1.8	59.3	577.0	
MB19	3.1	0.40	30.0	4.1	60.0	17.0	293.8	2.70	58.6	4.1	1.5	43.6	307.0	
MB21	5.8	0.33	26.0	4.7	48.0	32.0	278.0	1.50	22.0	4.0	1.1	35.3	181.0	
MB22	6.8	0.33	25.0	2.8	124.4	41.0	276.6	3.00	516.1	2.2	1.1	32.7	300.0	
MB23	8.9	0.64	16.6	2.8	20.6	23.0	231.0	2.50	17.4	7.7	1.2	25.9	276.0	
MB24	2.4	0.01	23.0	4.5	20.0	12.0	270.0	0.50	17.5	3.0	0.7	32.5	110.0	
MB25	1.7	0.04	6.4	1.6	6.0	3.0	88.7	0.16	4.8	1.2	0.3	10.1	46.0	
MB27	3.8	0.01	22.1	3.9	60.0	13.0	225.6	0.73	19.5	3.0	0.6	25.6	162.0	
MB28	7.1	0.01	20.2	4.2	13.0	10.0	227.3	0.33	11.8	2.9	0.6	18.5	151.0	
MB29	3.1	0.07	50.7	3.9	20.0	11.0	238.9	0.53	69.6	2.8	0.7	29.2	158.0	
MB30	2.0	0.01	3.9	0.8	24.0	5.0	80.9	0.15	1.9	0.4	0.2	8.2	49.0	
MB33	0.0	0.01	13.5	2.9	6.5	30.0	177.1	0.32	3.1	55.4	6.7	20.5	80.0	
MB34	3.6	0.01	27.5	5.8	16.0	11.0	386.3	0.49	9.0	5.3	1.1	43.5	87.0	
MB35	3.6	0.01	22.0	4.4	10.4	9.0	259.6	0.72	13.6	3.6	0.7	29.6	74.0	
MB36	1.0	0.01	17.9	4.0	10.6	7.0	292.6	0.43	8.0	3.6	0.8	30.7	75.0	
MB38	2.8	0.01	32.2	6.1	26.2	11.0	272.2	0.83	14.0	4.1	1.1	51.2	100.0	
MB41	4.0	1.20	15.2	2.4	100.0	77.0	105.2	0.76	10.0	0.6	0.5	10.0	3.013.0	
MB42	23.8	3.40	97.3	8.5	78.1	210.0	858.1	1.40	36.2	5.6	1.3	63.9	6.425.0	
MB43	3.1	0.84	3.6	0.6	14.4	38.0	37.2	1.10	0.6	0.6	0.2	4.8	959.0	
MB44	11.0	2.70	38.2	4.1	558	112.0	346.4	1.10	12.9	27.6	1.0	24.1	4,089.0	
MB45	6.1	1.40	16.0	2.1	594	52.0	83.5	1.10	9.1	2.1	0.8	20.3	367.0	
MB48	1.4	0.76	8.2	0.6	211	10.0	35.2	0.17	1.1	0.6	0.1	7.6	345.0	
MB47	2.0	0.12	4.2	0.6	39.1	6.0	43.4	0.36	0.6	0.6	0.2	7.3	67.0	

TABLE 2

trol was, however, observed for Cu, Pb and Zn which yielded weak (or negative) statistical correlations with Al_2O_3 , SiO_2 and TiO_2 (Pearson R < 0.2) and strong within-group covariability (R = 0.67 to 0.97). Sporadic enrichment of these heavy metals (two standard deviations greater than the mean) was recorded in the lower reaches of Port Kilindini (maxima 1177 mg kg $^{-1}$ Cu, 427 mg kg⁻¹ Pb, 283 mg kg⁻¹ Zn), Makupa Creek $(\text{maxima 43 mg kg}^{-1} \text{Cu}, 44 \text{ mg kg}^{-1} \text{Pb}, 225 \text{ mg kg}^{-1})$ Zn), the north-eastern shore of Port Reitz (maxima $26 \text{ mg kg}^{-1} \text{ Cu}$, $30 \text{ mg kg}^{-1} \text{ Pb}$, $190 \text{ mg kg}^{-1} \text{ Zn}$) and on the east of Mombasa Island. The anomalies can be tentatively ascribed to specific point sources including Kipevu oil terminal and the adjacent sewage outfall into Port Reitz; Kibarani landfill (Makupa Creek); and sewage outfalls from Mombasa Island into Port Kilindini and Mombasa Harbour.

Chemostratigraphic analyses of Mombasa cores highlighted enrichment (to two-fold greater than basal concentrations) of several heavy metals including Co, Cr, Cu, Ni, Pb, V and Zn in the uppermost 5–10 cm of sediment at most sites (Fig. 2). At Makupa Creek, a deeper though probably synchronous increase was observed at ca 20 cm depth, due to localized catchment disturbance (probably by landfill activities) and a resultant increase of sediment flux (Rees et al., 1996). With an average present-day sedimentation rate of 1- 2 mm yr^{-1} (Rees *et al.*, 1996), the accelerated sedimentary incorporation of potentially toxic metals may be broadly synchronous with urban development. There is, however, strong evidence that the profiles are controlled by a progressive transition from a quartzofeldspathic silt/sand assemblage with marine-derived shelly fragments (downward of 20 cm) to fine clastic material containing ca 20% organic matter in the surficial sediment. At most stations, the downcore profiles of Co, Cr, Cu, Ni, Pb and Zn were closely covariant with Al_2O_3 and TiO_2 (R > 0.7), and normalization against these major oxides effectively removed the concentration gradients portrayed in the bulk downcore plots.

Solid phase partitioning data for Makupa Creek (MB23) and Port Tudor (MB5) sediments confirmed that exchangeable (highly bioavailable) metals are of negligible significance to the total sediment load. Organo-metallic complexes form an important host

	Zr	585.0	285.0	185.0	163.0	259.0	312.0	641.0	224.0	839.0	304.0	0.721	726.0	164.0	162.0	233.0	208.0	188.0	163.0	0.611	0.921	234.0	178.0	498.0	401.0	222.0	200.0	253.0	503.0	460.0	272.0	383.0	383.0	535.0	120.0	90.0	1.0	173.0	321.0	518.0
	Zn	47.0	0.10	96.0	85.0	81.0	51.0	63.0	52.0	56.0	68.0 10	49.0	0. 12 0	139.0	239.0	283.0	86.0	118.0	225.0	156.0	100.0	/8.0	115.0	190.0	69.0	82.0	88.0	123.0	44.0	25.0	61.0	72.0	42.0	44.0	6.0	5.0	3.0	9.0 0.1	0.122	28.0
	v	0.69 0.00	00.0 143.0	163.0	139.0	108.0	46.0	104.0	101.0	72.0	96.0	0.6/	100.0	84.0	77.0	58.0	70.0	73.0	110.0	106.0	92.0	85.0	0.00	58.0	77.0	98.0	103.0	0.68	74.0	42.0	84.0	115.0	66.0	56.0	7.0	3.0	3.0	10.0	0.01	33.0
ons.	Sn	1.5	1.5	1.5	3.0	1.5	3.0	1.5	1.5	1.5	3.0	<u>, 4</u>		3.0	8.0	6.0	1.5	5.0	5.0	1.5	4.0	0.4	0.4	1.5	1.5	3.0	3.0	0.5	1.5	1.5	1.5	1.5	1.5	1.5	1.5	1.5	1.5	1.5	40	1.5
oling stati	Sr	107.0	75.0	80.0	105.0	358.0	1389.0	215.0	250.0	337.0	231.0	0.146	2110	510.0	562.0	816.0	690.0	639.0	170.0	151.0	168.0	290.0	0.104	151.0	236.0	160.0	149.0	217.0	137.0	138.0	117.0	93.0	122.0	293.0	2710.0	4118.0	2384.0	3802.0	0.175	685.0
asa samj	Ņ	16.0 20.0	34.0	38.0	32.0	28.0	12.0	23.0	16.0	15.0	24.0	14.0	0.01 010	23.0	26.0	17.0	21.0	22.0	30.0	30.0	27.0	23.0	0.42	17.0	21.0	26.0	29.0	24.0	17.0	10.0	21.0	29.0	14.0	15.0	3.0	1.0	1.0	2.0	3.0	8.0
т Моть	Pb	19.0	25.0	24.0	25.0	21.0	14.0	22.0	20.0	24.0	24.0	15.0	18.0	63.0	102.0	427.0	28.0	51.0	44.0	35.0	30.0	24.0	41.0	28.0	20.0	22.0	25.0	29.0	16.0	15.0	18.0	17.0	17.0	15.0	4.0	1.0	1.0	2.0	0.17	19.0
nents froi	Сп	15.0	31.0	41.0	26.0	24.0	12.0	19.0	11.0	16.0	19.0	0.71	16.0	39.0	70.0	177.0	24.0	30.0	43.0	33.0	26.0	21.0	30.05	16.0	17.0	23.0	24.0	27.0	13.0	7.0	17.0	26.0	12.0	13.0	3.0	4.0	1.0	5.0	0.04 0.0	8.0
TABLE 3 y weight) concentrations in interfacial sedime	Co	11.0	20.0	26.0	20.0	17.0	10.0	15.0	17.0	11.0	15.0	0.11	18.0	13.0	13.0	13.0 1	13.0	11.0	17.0	17.0	18.0	14.0	160	12.0	16.0	17.0	18.0	0.61	13.0	8.0	16.0	20.0	12.0	13.0	0.5	1.0	0.5	2.0	0.4 0	5.0
	Cr	45.0 54.0	0.4.0	105.0	94.0	71.0	26.0	63.0	66.0	50.0	63.0	43.0	46.0 65 0	59.0	61.0	64.0	50.0	57.0	81.0	70.0	66.U	0.73	0.10	52.0	51.0	65.0	72.0	63.0 57.0	51.0	25.0	60.0	82.0	47.0	46.0	4.0	0.5	0.5	4.0	0.04	19.0
	Ba	540.0	333.0	349.0	268.0	287.0	268.0	365.0	312.0	448.0	408.0	55/.U	407.0	555.0	359.0	457.0	299.0	304.0	322.0	251.0	309.0	369.0	326.0	373.0	357.0	324.0	336.0	379.0	626.0	660.0	412.0	334.0	524.0	374.0	139.0	44.0	34.0	71.0	041.0	459.0
	As	10.0	0.6	6.0	18.0	18.0	16.0	15.0	12.0	10.0	14.0	0.01	15.0	15.0	13.0	13.0	14.0	12.0	11.0	13.0	12.0	0.11	0.01	5.0	8.0	13.0	11.0	0.0	5.0	2.0	7.0	8.0	4.0	10.0	1.0	1.0	1.0	2.0	0.0	4.0
	TiO ₂	0.585	1.077	1.223	1.039	0.819	0.370	0.785	0.774	0.811	0.749	0.480	200-0 0 771	0.689	0.679	0.527	0.600	0.627	0.837	0.736	0.784	0./30	0.802	0.559	0.653	0.810	0.876	0.757	0.634	0.424	0.761	0.931	0.569	0.556	0.039	0.026	0.008	0.094	0.000	0.362
g kg ⁻¹ d	Na ₂ O	2.5	2.6 2.6	3.3	4.8	3.5	2.0	2.8	4.7	2.4	3.1	0.0	- 0 - 0	5.4	7.6	2.9	2.6	3.2	4.1	10.0	10.0	3.1	4.0 4 4	1.7	7.1	7.5	6.0 2	3.0 2 2	2.6	2.8	2.9	4.1	2.5	1.9	1.9	1.9	2.2	2.2	4.0	2.1
ment (m _g	SiO ₂	64.46 67 17	02.12 51.81	49.77	46.48	47.92	45.68	53.10	47.58	57.86	50.88	41.22	15 65	43.68	41.59	37.86	46.48	42.05	47.90	38.32	37.96	10.84	47.07	54.12	49.66	46.39	48.84	49.48 52.70	57.07	63.30	52.47	50.55	58.19	55.90	28.36	5.91	2.93	16.76	04.00 21 15	55.09
trace ele	K ₂ O	2.46 7.86	2.59	2.56	2.48	2.37	1.52	2.44	2.24	2.30	2.51	1.8/	07 7	2.22	2.15	1.80	2.14	2.12	2.44	3.33	2.06	2.37	27.2 2 76	1.98	2.12	2.29	2.48	2.34	2.31	2.26	2.40	2.56	2.28	2.06	0.83	0.23	0.19	0.55	74 177	2.15
(%) and	MnO	0.031	07070	0.074	0.061	0.032	0.023	0.030	0.035	0.033	0.057	0.040	100.0	0.043	0.044	0.042	0.055	0.040	0.031	0.036	0.043	0.042	70.046	0.029	0.037	0.047	0.046	0.039	0.051	0.017	0.034	0.029	0.035	0.030	0.006	0.005	0.005	0.006	0.037	0.014
jor oxide	MgO	1.3 1 5	2.0	2.1	2.3	2.1	1.2	1.9	2.0	1.5	2.1	8.1	- 1.4 0 0	2.6	2.4	2.4	2.3	2.5	2.5	2.2	1.7	2.7	9 r 7 r	1.5	2.1	2.4	2.6	5.7	9	1.0	2.1	2.4	1.5	1.9	2.2	2.1	3.0	2.0	7.1	1.4
Ma	Fe ₂ O ₃	3.88	7.31	8.34	7.37	5.92	3.14	5.34	5.84	4.10	5.06	40.4 2003	07.0 2 70	5.19	5.19	4.49	4.34	4.59	6.35	5.81	5.80	/0.c	20.0	3.26	4.57	5.83	6.15	5.16	4.19	2.33	4.90	6.40	3.64	3.96	0.17	0.07	0.06	0.28	2.04 0.80	1.59
	CaO	0.59	0.44	0.36	0.94	4.41	16.25	3.39	6.37	4.66	2.60	13.92	00 6	6.73	7.83	14.61	9.15	8.20	2.14	1.67	1.67	4.5 4.5	70.0 70 E	5.71	2.85	1.63	1.49	3.73	1.42	0.34	0.97	0.85	0.66	4.19	32.79	44.34	44.84	38.58	17.1	8.90
	Al ₂ O ₃	10.94	16.44	16.44	15.43	13.72	6.87	12.97	12.53	10.23	13.66 2.57	9.70	10.03	14.41	13.40	9.79	12.43	12.88	15.18	12.58	12.20	14.80	14.50	11.51	13.26	14.66	16.91	15.07	13.58	11.49	15.85	15.83	13.37	10.89	1.22	0.32	0.23	1.06	00./ 1.8.1	7.22
	Sample	MBI	MB3	MB4	MB5	MB6	MB7	MB8	MB9	MB10	MBII	MBIZ	MR14	MB16	MB17	MB18	MB20	MB21	MB22	MB23	MB24	MB25	MB28	MB29	MB30	MB31	MB 32	MB33	MB35	MB36	MB37	MB38	MB39	MB40	MB41	MB42	MB43	MB44	MB46	MB47

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Fig. 2 Typical downcore trace metal profiles (mg kg⁻¹ dry weight) in inshore sediments around Mombasa, exemplified by data from Tudor Creek site MB1.

for Cu (up to 50% of the total), V (up to 25%) and Zn (up to 10%), with fulvic fractions systematically dominant (ca 80% of the total organic loading). hydrochloride-extractable phases Hydroxylamine (mainly hydrous Mn oxides) constitute the major carrier of Mn (>50% of the total), Co (50% total), Cu (30% total), Pb (100% total) and Zn (ca 60% total) in the uppermost 60 cm of sediment. These oxides were probably precipitated under aerobic conditions at the sediment-water interface, and are subject to progressive (kinetically constrained) dissolution during burial. Data acquired from the aqua-regia stage of the leaching sequence indicated that aluminosilicates, detrital sulphides and other resistate minerals dominated with respect to the total budgets of Al (>85%), Cr (>85%), Fe (>60%), Ni (>70%) and V (>50%) at all stratigraphic levels.

GCMS analyses of surficial sediments from Makupa Creek (MB23), Port Reitz (MB27) and Port Tudor (MB1) yielded significant anomalies at the former site for toluene (12 mg kg⁻¹ dry weight), hexanone (21 mg kg⁻¹), *n*-alkanes (24 mg kg⁻¹) and a complex suite of heavy hydrocarbons (3600 mg kg⁻¹). The concentrations of 16 PAH compounds in sediments from these localities were universally low (<1 mg kg⁻¹). With the exception of a single anomaly for DDT

(0.46 mg kg⁻¹) at a Tudor Creek site (MB5), all surficial sediments yielded values of <0.1 mg kg⁻¹ dry weight for major organochlorine groups including DDE, DDT, dieldrin, α -endosulphan, β -endosulphan, endrin, α -HCH, β -HCH hexachlorobenzene and trifluralin.

In view of the magnitude of annual industrial and domestic discharges currently known to prevail in the region the strictly limited evidence of degeneration of inshore/nearshore sediment and water quality attributable to Mombasa's recent urban expansion is striking. In addition to the suspended solid influxes derived from sewage (>3500 t yr⁻¹), stormwater runoff $(>9000 \text{ t yr}^{-1})$ and routine industrial activities $(>20\ 000\ t\ yr^{-1};\ Munga\ et\ al.,\ 1994),\ the\ Mombasa$ creeks have been subject to at least two significant oilspills in the past decade, the geochemical effects of which are now difficult to detect. This can be ascribed to specific physical, hydrodynamic and biogeochemical attributes which promote effective contaminant dispersal and/or degradation, as detailed by Williams et al. (1996a).

The LOCS Mombasa survey has provided a comprehensive baseline against which the effects of future urban and industrial expansion can be evaluated. The establishment of an appropriate system for

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integrating the LOCS datasets with a wider range of environmental and socio-economic information is, however, vital to maximize their practical utility in coastal zone management (CZM). To this end, the LOCS database has been downloaded to a pre-existing thematic GIS for the Kenya coast, developed by UNEP OCA/PAC under a sub-project (EAF/6) of the Action Plan for the Protection, Management and Development of the Marine and Coastal Environment of the East African Region (Williams *et al.*, 1996b). It is anticipated that this integrated system will ultimately serve as a model for adoption throughout the East African region.

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Imposex in Three Species of *Thais* from Singapore, with Additional Observations on *T. clavigera* (Küster) from Japan

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Observations on imposex are provided for Thais bitubercularis, T. clavigera and T. jubilaea (Mollusca: Neogastropoda: Muricidae) from Singapore. The results suggest that the three species are affected by tributyltin. Thais bitubercularis appear to be least affected, while T. jubilaea was the most affected, in terms of the size of the pseudopenis. The sex ratios in most cases were in favour of the males, suggesting a higher mortality amongst females. Relative penial size (RPS) values were less than 25% but aborted capsules were present in the lumen of the capsule glands of females examined. In contrast, additional observations of T. clavigera from Japan had RPS values up to 88%. These differences probably reflect seasonal variation in male penial size, but the possibility that females respond to high concentrations of TBT by growing a larger pseudopenis despite blockage of the oviduct, should not be discounted. © 1997 Elsevier Science Ltd

It has previously been established that tributyltin (TBT) causes the imposition of male sex characters in females of neogastropods, and this is particularly well-reported in gastropods belonging to the family Muricidae (Mollusca: Prosobranchia; see Gibbs and Bryan, 1986; Bryan et al., 1988; Fioroni et al., 1992). The morphological characteristics induced in females exposed to TBT have been used to plot a dose-response relationship for some temperate species of the genera Nucella and Ocenebra (Gibbs et al., 1987, 1990). At low concentrations (0.5–1 ng TBT 1^{-1}) the females grow a small non-functional penis, and an incomplete vas deferens. At high TBT concentrations (typically above 5 ng TBT 1^{-1}) the females have a fully grown penis and vas deferens comparable in size with those of mature males (Gibbs and Bryan, 1986). In some species, the vas deferens obstructs the vulva, causing a blockage of the oviduct (e.g. Gibbs et al., 1987; Oehlmann et al., 1991). This prevents the females from laying eggs, often resulting in the accumulation of eggs in the oviduct, and the eventual rupture of the capsule gland (Gibbs and Bryan, 1986; Gibbs et al., 1987). The presence of other male sex organs, such as the prostate, seminal vesicles and testes, has been found in some females affected by high TBT concentrations (Gibbs et al., 1988) although in a few populations affected by TBT, aphally has been reported in some individuals (Gibbs, 1993; Huet et al., 1996). Effects are also felt at the population level. Because the young of Nucella lapillus undergo direct development (i.e. the eggs develop into miniature adults without passing through a planktonic stage) and have limited dispersal capabilities, populations rendered sterile by TBT become extinct, since there is little or no immigration (Bryan et al., 1986).

Studies on species of Muricidae in the Malayan region are hampered by taxonomic confusion, particu-